

pubs.acs.org/est

Article

Partitioning and Mobility of Chromium in Iron-Rich Laterites from an Optimized Sequential Extraction Procedure

Ruth Esther G. Delina,* Jeffrey Paulo H. Perez, Jessica A. Stammeier, Elena F. Bazarkina, and Liane G. Benning



Philippines. Our results show that other SEPs inadequately leach Cr host phases and underestimate the Cr fractions. Our SEP recovered up to seven times higher Cr contents because it (a) more efficiently dissolves metal-substituted Fe phases, (b) quantitatively extracts adsorbed Cr, and (c) prevents overestimation of organic Cr in laterites. With this new SEP, we can estimate the mineral-specific Cr fractionation in Fe-rich tropical soils more quantitatively and thus improve our knowledge of the potential environmental impacts of Cr from lateritic areas.

KEYWORDS: Cr(VI), dissolution, iron (oxyhydr)oxides, metal substitution, mineral synthesis, nickel laterite, SEP optimization, SEP validation

INTRODUCTION

Laterite broadly refers to the iron (Fe) or aluminum (Al) (oxyhydr)oxide-rich weathering mantle covering about 33% of the continents.^{1,2} Laterites developed from tropical weathering of ultramafic rocks (e.g., peridotites and dunites) predominantly consist of Fe (oxyhydr)oxides in the form of goethite (FeOOH) and hematite (Fe₂O₃) and are often enriched with critical metals such as nickel (Ni), cobalt (Co), and scandium (Sc) mainly incorporated into the minerals.^{3–5} Such metal deposits are known as nickel laterites, and they are the world's main source of Ni, accounting for ~60% of the global production.^{4,6}

efficiency of our new SEP through analyses of laterites from the

Nickel laterites also contain elevated concentrations of chromium (up to ~70,000 mg kg⁻¹)^{7,8} that are multiple orders of magnitude higher than upper crustal averages (35 mg kg⁻¹).⁹ Chromium commonly occurs as Cr(III) and Cr(VI), with the latter being a highly mobile, toxic, and carcinogenic pollutant.^{10–12} The majority of Cr in Ni laterites is present as Cr(III) and preferentially substitutes into octahedral sites of Fe (oxyhydr)oxides (e.g., goethite and hematite), silicates, and spinels,^{7,13} while Cr(VI) predominantly exists as oxyanions (i.e., HCrO₄⁻ and CrO₄²⁻) adsorbed onto these same

minerals^{14,15} or dissolved in pore waters or soil solutions.¹⁶ Since Ni laterites are exploited through large opencast surface mining, Cr(VI) leaches into surface- and groundwaters where it can reach concentrations (up to 1600 μ g L⁻¹),¹⁷⁻¹⁹ far exceeding international drinking water standards (50–100 μ g L⁻¹).^{20,21} Elevated levels of Cr(VI) in these water resources can lead to, so far, not well-understood health issues for the local population.⁷ Thus, it is important to quantify the partitioning and possible transport mechanisms of Cr species in laterite host phases and evaluate how these Cr-mineral phase specific associations dictate the potential mobility, bioavailability, and toxicity of Cr during geogenic (e.g., weathering) and anthropogenic (e.g., mining) processes.

The partitioning of elements in soils and sediments is traditionally evaluated through sequential extraction proce-

Received:December 20, 2023Revised:March 18, 2024Accepted:March 18, 2024Published:March 29, 2024





extractant		ferrihydrite		goethite		hematite		magnetite (synthetic)		magnetite (natural)		chromite
	poorly cryst. Fe Ox		Ox	crystalline Fe Ox								residual
	Fe dissolution efficiency (%)										_	
0.1 M Ca(NO ₃) ₂	bdl	!		bdl	ŀ	dl		bdl		bdl		bdl
0.01 M NH ₄ H ₂ PO ₄	0.0	$2 (8 \times 10^{-1})$	⁴)	bdl	k	odl		bdl		bdl		bdl
1 M NaOAc	1.40 (0.06)			bdl bdl		dl	ba			$0.15~(6 \times 10^{-3})$		bdl
0.1 M NH ₂ OH·HCl	$0.27~(6 \times 10^{-3})$			<i>bdl</i> 0.0		$0.06 (1 \times 10^{-3})$	$5(1 \times 10^{-3})$ 0		$\times 10^{-3}$)	bdl		bdl
5% NaOCl (1:20, 2×)	$0.004 (1 \times 10^{-4})$			bdl bdl		odl	l			bdl		bdl
1 M HCl, 8 h	99.	99.2 (2.3)		0.51 (0.02) 15		5.4 (0.5)	4 (0.5)		9.52 (0.32)		$0.14 (3 \times 10^{-3})$	
6 M HCl, 75 °C, 24 h	NI	ND		104 (3) 105		05 (4)	(4) 99.0 (3		97.6 (2.6)		.6)	bdl
extractant	Cr(VI)	-ads FHY	Cr(VI)-a	nds Goe	Cr(VI)-ad	ds Hem	Cr-FHY	7	Cr-Goe	Cr-He	m	chromite
	ads		adso	rbed	d		orly cryst. Fe Ox		crysta	crystalline Fe Ox		residual
	Cr dissolution efficiency (%)											
$0.1 \text{ M Ca}(\text{NO}_3)_2$	1.20	(0.06)	6.12 (0.31)	7.04 (0	0.36)	ND		ND	ND	N	D
0.01 M NH ₄ H ₂ PO ₄	76.8	(3.9)	81.3 (4.2)	75.9 (3	3.9)	ND		ND	ND	Ν	D
5% NaOCl (1:5)	ND		ND		ND		4.07 (0.0	9)	49.2 (1.4)	4.76 (0.	.08) 0.	$15 (6 \times 10^{-4})$
5% NaOCl (1:20, 2×)	ND		ND		ND		26.0 (1.4)	60.5 (1.7)	15.9 (0.	.4) 0.	$12 (1 \times 10^{-3})$
0.5 M HCl, 4 h	ND		ND		ND		80.6 (2.1)	ND	ND	Ν	D
1 M HCl, 4 h	ND		ND		ND		96.4 (2.5)	ND	ND	Ν	D
1 M HCl, 8 h	ND		ND		ND		98.4 (2.5)	11.6 (0.6)	2.14 (0.	.07) ba	11
6 M HCl, 50 °C, 48 h	ND		ND		ND		ND		52.8 (2.2)	98.6 (5.	.6) bc	11
6 M HCl, 75 °C, 24 h	ND		ND		ND		ND		94.7 (5.7)	102 (6)	ba	11
extractant		mixture 1 mixture 2 mixture			mixture 3	3 OREAS 182			OREAS 190			
		Cr d	issolution	efficiency	(%)	Cr ext.	(%)	Fe ex	t. (%)	Cr ext. (9	%)	Fe ext. (%)
step 1: 0.1 M Ca(NO ₃) ₂						0.36 (0.01)	bdl		$0.04 (1 \times 1)$	0^{-3}) ba	1 1
step 2: 0.01 M NH ₄ H ₂ PO ₄		51.2 (0.1)				0.47 (0.01)	0.45 (0.	.01)	$0.12 (3 \times 1)$	0^{-3}) 0.	$26 (6 \times 10^{-3})$
step 3: 1 M NaOAc						0.01 (2 ×	10^{-4})	0.40 (0.	.01)	0.56 (0.02)	0.	$37 (8 \times 10^{-3})$
step 4: 0.1 M NH ₂ OH·HCl						0.03 (1 ×	10^{-3})	0.23 (5	$\times 10^{-3}$)	$0.03 (1 \times 1)$	0^{-3}) 0.	$18 (4 \times 10^{-3})$
step 5: 1 M HCl, 8 h			102	(2)	89.9 (1.9)	1.01 (0.03	3)	5.05 (0.	.11)	0.80 (0.02)	3.	94 (0.08)
step 6: 6 M HCl, 75 °C, 24 h		92.0 (1.9)	89.1	(1.9)	108 (2)	11.0 (0.3)	1	82.7 (1	.8)	14.8 (0.4)	7:	5.5 (1.6)
step 7: 5% NaOCl (1:20, 2×) ^b						1.26 (0.04	+)	bdl		0.43 (0.01)	ba	dl
residual		106	109		101							

Table 1. Dissolution Efficiencies of Single Extractions on Selected Mineral Standards and Sequential Extractions on Mixtures and Laterite CRMs^a

^{*a*}FHY—ferrihydrite, Goe—goethite, Hem—hematite, Ox—(oxyhydr)oxides. *Note:* dissolution efficiency or metal extracted (ext.) (%) = (wt % extracted/wt % total) × 100. (#)—analytical uncertainty (<5% relative) based on multiple measurements ($n \ge 5$) of QC solutions. Mixture compositions are further detailed in Table S5. For mineral mixtures, the residual fraction dissolution efficiency was represented by the (wt % total Cr – \sum wt % non-residual)/wt % Cr in chromite. *bdl*—below detection limit; ND—no data. ^{*b*}The NaOCl step (step 7) was applied after the 6 M HCl treatment (step 6) to prevent the indiscriminate oxidation of Cr from Fe (oxyhydr)oxides.

dures (SEPs), which are based on a series of increasingly aggressive reagents that categorize the leached elements into chemical or mineralogical fractions.^{22,23} However, SEPs are criticized for poor selectivity of extraction reagents, redistribution of metals, and incomplete dissolution.^{22,24-26} For instance, in the case of Cr, most SEPs cannot completely dissolve common host phases such as chromite and Fe (oxyhydr)oxides.^{18,27} Chromites are highly recalcitrant to dissolution with most conventional digestion methods, 18,24,28 while the dissolution of Fe (oxyhydr)oxide is known to be affected by metal substitution. For example, substitution of Cr and Al for Fe in goethite has been shown to strongly inhibit its dissolution in strong acids and reductants;^{23,29} yet the effect of metal substitution is often overlooked when developing SEPs. In addition, SEPs are commonly optimized for cationic species,²² and thus when applied to Cr, they likely underestimate the distribution of Cr(VI) oxyanions. More importantly, no SEP has been critically assessed for its suitability for Cr partitioning in tropical laterites, which possess such a unique Fe mineral assemblage. Existing SEPs applied for Cr fractionation in tropical soils rich in Fe and Mn

(oxyhydr)oxides^{18,28,30,31} were originally developed for other metals and/or sample matrices. These include the modified Geological Survey of Canada (mGSC) procedure, which was initially developed to partition Cd in temperate soils³² but has also been tested to be suitable for tropical soils.³³ The SEP used in Quantin et al.²⁸ was adapted from well-cited procedures, including Tessier et al.,³⁴ which were intended for extracting metals such as Si, Ca, Cd, and Fe from river sediments and temperate to subtropical soils.^{34–36} Finally, the SEP by Silveira et al.²³ was designed for tropical soils and optimized for Zn, Cu, Fe, and Mn but not Cr. Because these SEPs are optimized neither for Fe-rich laterites nor for Cr species, there is a need to optimize a Cr SEP and thus provide a more quantitative evaluation of the fate and potential impacts that Cr can have in such lateritic environments.

To address this gap, we have characterized the partitioning of Cr in various tropical Ni laterite profiles using a new SEP for Fe (oxyhydr)oxide-rich laterites. We optimized different extractants using Cr- and Fe-bearing phases commonly present in Ni laterites and certified laterite references and validated our new SEP using Ni laterites from different localities in the Philippines. We also compared and contrasted our results with the three aforementioned SEPs^{23,28,32} and documented the far more efficient and targeted nature of our new SEP.

MATERIALS AND METHODS

Natural Laterites. The partitioning of Cr was examined in previously well-characterized Ni laterites from three major Ni mining districts in the Philippines (Palawan,¹⁸ Zambales, and Surigao³⁷). Palawan samples described in Delina et al.¹⁸ were obtained from a 6.8 m thick Ni laterite profile consisting of an upper Fe (oxyhydr)oxide (i.e., goethite and hematite) dominated limonite zone and a lower silicate-rich (i.e., serpentine and smectite) saprolite layer separated by a thin transition zone. From bottom to top, the profile is characterized by a dramatic increase in Cr and Fe contents (from 0.5 to 2.9 wt % Cr and 9 to 54 wt % Fe).¹⁸ Samples from the limonite, transition, and saprolite zones (hereafter referred to as PAL-1, PAL-2, and PAL-3, respectively), representative of different Cr and Fe concentrations, were used to evaluate the efficiency of our new SEP. Furthermore, the robustness of our SEP was tested on five high Cr (1.1–1.7 wt %) and Fe (38–55 wt %) limonite samples from Zambales (ZAM-1 to ZAM-3) and Surigao (SUR-1 to SUR-2). These primarily contain goethite (>89%) with minor spinel (2.4-11%). Characterization of these samples is discussed in Text S1.

Synthesis and Preparation of Mineral Standards. Various synthetic and natural mineral references (Table S1) representing the composition of Fe (oxyhydr)oxide-rich laterites^{3,18,38} were prepared to optimize the SEP. Pure and metal (Me)-substituted (Me = Al and Cr) ferrihydrite, goethite (α -FeOOH), hematite (α -Fe₂O₃), and pure magnetite [Fe-(II)Fe(III)₂O₄] were synthesized using standard procedures adapted from Schwertmann and Cornell.³⁹ In addition to Cr, Al-substituted Fe minerals were also prepared since pedogenic Fe (oxyhydr)oxides often structurally incorporate Al.²⁹ Cr-(VI)-adsorbed Fe (oxyhydr)oxides were also prepared. Details of the preparation and characterization of these synthetic minerals and natural samples (e.g., chromite) can be found in Text S2.

Testing and Optimization Based on the Mineral References. Single extractions (detailed in Text S4) were carried out to assess the dissolution efficiency and selectivity of different reagents. Selection of extractants were based on extensive reviews of SEPs^{22,25,40} and procedures applied to Fe (oxyhydr)oxides and Fe-rich soils and sediments.^{23,41-43} Operating conditions (e.g., temperature, duration, and solid-to-liquid ratio) and concentrations were varied and tested to find the best possible extractant (Tables 1 and S3).

To partition adsorbed Cr(VI) oxyanions, we applied an alkaline (pH 8) 0.01 M NH₄H₂PO₄ treatment for 16 h^{44,45} (see Text S3 for detailed information) on Cr(VI)-adsorbed Fe (oxyhydr)oxides. We evaluated the selectivity of typically used extractants for the prior exchangeable fraction step [i.e., 0.1–1 M Ca(NO₃)₂ and 1 M MgCl₂ for 2 h]²⁵ with respect to the Cr(VI)-adsorbed phases. We also examined the effect of the following treatments on Cr- and Fe-bearing minerals: 1 M NaOAc buffer (pH 4.5) for 5 h (carbonate-bound fraction),^{35,41} ~5% NaOCl (pH 8.5) at boiling temperature for 30 min (organic fraction),^{46,47} and 0.1 M NH₂OH-HCl in 0.01 M HNO₃ for 10 min (Mn phase-bound fraction).⁴⁸ Furthermore, we assessed the effectiveness of different concentrations of HCl in dissolving Fe (oxyhydr)oxides of different crystallinities. We tested dilute (0.5 and 1 M)

 $\mathrm{HCl}^{42,49}$ for the poorly crystalline fraction and 6 M $\mathrm{HCl}^{23,50}$ extractions at different temperatures (50 and 75 °C) and reaction times (\leq 48 h) for the crystalline fraction.

Sequential Extractions. Based on the single extractions, we optimized a new SEP and tested it on mixtures of mineral references (Table S5) and Ni laterite certified reference materials (CRMs) (OREAS 182 and 190). The optimized SEP was applied to the Ni laterites and compared to the three SEPs previously used for Cr partitioning in laterites and related tropical soils: the mGSC procedure³² (SEP 1), the SEPs used in Quantin et al.²⁸ (SEP 2), and Silveira et al.²³ (SEP 3) (outlined in Table S2). All SEPs were performed in duplicate on the Palawan Ni laterite samples except for SEP 1, which was previously applied to the same samples in Delina et al.¹⁸

In each SEP step, reagents were mixed with powdered samples in acid-cleaned centrifuge tubes and reacted in temperature-controlled orbital shakers at 150–250 rpm. Liquid phases were separated from the residue by centrifugation at 10,052g for 10 min. Between each extraction, residues were washed with Milli-Q water (~18.2 M Ω ·cm) and freeze-dried before the next extraction step. The analysis of the supernatants was identical to that of the single extractions (Text S5). Relative standard deviations (RSDs) of Cr were <5% for ~80% of samples. RSD >10% was observed in extracts with Cr concentrations near the quantification limit.

The residues of the sequential extraction are chemically resistant minerals such as chromite (see Figure 4) that are highly prone to incomplete dissolution by conventional digestion methods.^{18,24,25} In our work, acid digestion post Na_2O_2 fusion⁵¹ (see Text S1) did not lead to full dissolution, and dark-colored chromite grains persisted. We accounted for the Cr associated with this residual fraction as the difference between the total concentration and the sum of all extracted non-residual fractions, and we mainly discussed and compared steps that target the latter.

To characterize the residual fraction and understand how different SEPs extract Cr, we analyzed the mineralogy and local bonding environment of Cr in selected SEP residues after the crystalline Fe phase-bound step using X-ray diffraction (XRD), scanning electron microscopy (SEM), and high-energy resolution fluorescence detection X-ray absorption spectroscopy (HERFD-XAS), as fully described in the Supporting Information.

RESULTS AND DISCUSSION

Metal Substitution in Natural Fe (Oxyhydr)oxides. XRD and infrared (IR) spectroscopy patterns of Fe (oxyhydr)oxide-dominated samples from each Ni laterite district (Figure 1) showed patterns consistent with those of the synthetic goethites (Figures S1 and S2). PAL-1, containing nearly equal amounts of goethite (48%) and hematite (43%), exhibited combined patterns of the Fe phases. Diffraction peaks of the natural samples showed remarkable shifts to higher angles or Q $(=2\pi/d)$ values compared with pure goethite $(\Delta Q_{(110)} \leq$ 0.012) and hematite ($\Delta Q_{(110)} = 0.003$) (Figure 1b,c), suggesting metal substitution. This is supported by the similar shifts displayed by substituted goethites ($\Delta Q_{(110)} \leq 0.009$) and hematites ($\Delta Q_{(110)} \leq 0.015$), indicating a decrease in unit cell volume due to the smaller octahedral radii of Al(III) (0.530 Å, 18% smaller) and Cr(III) (0.615 Å, 5% smaller) compared to Fe(III) (0.645 Å). Consistent with previous studies,⁵²⁻⁵⁵ Alsubstituted phases showed larger shifts due to the significantly smaller atomic radius of Al. The effect of substitution was also



Figure 1. XRD patterns and IR spectra of the Fe (oxyhydr)oxide-rich laterites compared with pure and metal-substituted Fe phases. (a) XRD patterns with highlighted (110) diffraction peaks of (b) goethite (Goe) and (c) hematite (Hem). (d) IR spectra of the samples with highlighted (e) OH bonds of goethite and (f) the Fe–O bond of hematite. Dashed vertical lines highlight peak shifts relative to pure phases.

observed in the IR spectra (Figure 1e,f), where the separation of the OH bending modes of synthetic goethites at \sim 790–890 cm⁻¹ increased from 95 to 103 cm⁻¹ and the Fe–O band of synthetic hematites at \sim 520 cm shifted to higher wavenumbers.

Among the natural goethites, PAL-1 showed the largest diffraction peak shift and IR band separation, which are slightly higher than those of the Al-goethite. This may suggest a higher extent of substitution of many different cations, bigger differences in the atomic radii of substituting metals, or crystal disorder.^{52,54–57} Aside from Al and Cr, Fe (oxyhyr)oxides in laterites have been found to be important hosts for Ni, Co, and Mn.^{55,58,59}

Cr and Fe Extractability from Mineral Standards. Given that Cr in laterites could occur as adsorbed species or structurally incorporated in predominant Fe (oxyhydr)oxides, and metal substitution could affect the crystal structure, and thus, the solubility and dissolution rate of these Fe phases,⁵⁹ we tested the efficiency and selectivity of different extraction steps with a range of Cr and Fe minerals (Tables 1 and S3).

Easily Mobilizable Fractions. Our data show that the 0.01 M $NH_4H_2PO_4$ treatment effectively desorbed more than 70% of Cr from the Cr(VI)-adsorbed Fe (oxyhydr)oxides, with negligible Fe dissolution. A disadvantage of phosphate treatment is that residual adsorbed phosphate can retard Fe dissolution^{49,60} by surface passivation, decreasing the reactivity of the Fe (oxyhydr)oxides.^{61–63} This was evident in the incomplete recovery of Fe from goethite and natural magnetite when the 6 M HCl extraction was preceded by phosphate

treatment (Table S4). It is therefore necessary to perform a rinsing step after phosphate extraction. While Ruttenberg⁶⁴ recommended MgCl₂ wash for phosphorus extractions, we decided to use ultrapure water to minimize dissolved salts in the extract and avoid possible interferences during measurements. A minimum of 3 successive water rinses were found sufficient to displace most of the phosphate (Figure S4).

 $Ca(NO_3)_2$ and MgCl₂ extractions for the exchangeable fraction, usually applied at the beginning of SEPs, indiscriminately extracted up to 20% of Cr from the Cr(VI)-adsorbed Fe (oxyhydr)oxides (Figure S5). To avoid substantial underestimation of adsorbed Cr, the most dilute $Ca(NO_3)_2$ (0.1 M) treatment that extracted only 1–7% of Cr was chosen for the exchangeable fraction. Overall, these experiments imply that previous SEPs without a phosphate step and using only nitrate or chloride salts underestimated the easily mobilizable Cr fraction (exchangeable and adsorbed).

Non-Fe Phase-Bound Fractions. Acetate, hypochlorite, and hydroxylamine hydrochloride extractions partitioned very little amounts of Fe (<1%) from the reference phases (Table 1). However, NaOCl leached a significant amount of Cr from all Cr-substituted Fe (oxyhydr)oxides. Among the extractants used for organic matter (e.g., NaOCl, H_2O_2 , and $Na_4P_2O_7$), NaOCl was reported to exhibit greater efficiency and minimal attack on amorphous Fe (oxyhydr)oxides and clays in soils.^{22,25,65} However, our results clearly showed that NaOCl treatment leads to substantial Cr release, irrespective of the S/ L ratio used. The typical 1:5 ratio²³ extracted 4-49% of the Cr incorporated in the Fe (oxyhydr)oxides; meanwhile, the 1:20 ratio performed once and twice^{46,47} released 12–56 and 16– 60%, respectively. Extracts showed faint to strong yellow hues, suggesting the presence of chromate, and hence, the possible oxidation of Cr(III) to Cr(VI). This aligns with the prior work on the oxidative dissolution of Cr(III) hydroxide with NaOCl.^{66,67} Earlier SEPs of other metals also reported the indiscriminate oxidation of redox-sensitive elements by NaOCl. Gruebel et al.⁶⁸ and Wright et al.⁶⁹ revealed that adsorbed and incorporated Se species in selenides were oxidized to Se(VI), leading to substantial overestimation of the organic pool. Similarly, La Force and Fendorf⁴⁶ showed that Fe(II) from mine wastes were oxidized by NaOCl, resulting in inaccurate partitioning of Fe. Therefore, to avoid the indiscriminate oxidation of Cr, we applied the NaOCl treatment after the Fe (oxyhydr)oxide dissolution.

Poorly Crystalline Fe Phase-Bound Fraction. Among the tests using 0.5 and 1 M concentrations and a duration of 4-8 h (Tables 1 and S3), the 8 h 1 M HCl extraction was found to be most effective for poorly crystalline Fe (oxyhydr)oxides. All ferrihydrites were dissolved with >97% total Fe recovery, with Cr-substituted ferrihydrite showing the least recovery. To test the selectivity of 1 M HCl, we applied it to crystalline phases. It extracted up to 15% of total Fe in pure synthetic phases, comparable with earlier works showing dissolution of up to 33% of Fe from synthetic hematite and 9% from magnetite.^{42,49} It should be noted that pure minerals, such as these, rarely occur in nature, and thus, the selectivity of reagents is better evaluated with respect to metal-substituted and natural phases. The substantially low Fe, Cr, and Al dissolution efficiencies (below detection to 12%) (Tables 1 and S3) from metalsubstituted and natural Fe (oxyhydr)oxides validate the selectivity of the 1 M HCl treatment.

Crystalline Fe Phase-Bound Fraction. We optimized a 6 M HCl extraction that has been used for the sequential extraction

of crystalline Fe oxides and/or sheet silicates.^{50,70-72} Extractions at 50 °C after decreasing the S/L ratio (1:40 to 1:100) and increasing the duration (24–48 h) compared to previous SEPs²³ did not completely dissolve the Fe (oxyhydr)-oxides, especially Cr-goethite. Full dissolution was only achieved after further increasing the temperature to 75 °C. Time series experiments (Figure 2) revealed that all crystalline Fe (oxyhydr)oxides, except for Cr-goethite, were effectively dissolved within 2 h. Cr-goethite was only fully dissolved after 24 h, while chromite was unaffected even after 48 h. In comparison to prior dissolution of goethites using 6 M HCl at 25 °C,²⁹ the optimized 6 M HCl treatment reduced the time to fully dissolve Al-substituted goethite from ~220 to 2 h and



Figure 2. Dissolution time curves of reference Fe minerals: (a) goethites (Goe), (b) hematites (Hem), and (c) spinels (Mag—magnetite, Chr—chromite) in 6 M HCl at 75 °C. Error bars indicate analytical uncertainty (<5% relative) based on multiple measurements (n = 8) of QC solutions.

increased the dissolution extent of Cr-goethite from <50% after 350 h to >98% after only 24 h.

Faster rates of dissolution were observed from pure minerals compared with their metal-substituted and natural counterparts. These findings are consistent with previous works on synthetic Fe (oxyhydr)oxides dissolved in HCl,^{29,73,74} where they attributed the slower dissolution rates of Cr- and Al-substituted Fe phases to the higher bond strength of the Me–OH/O bonds [e.g., Al(III)–O = 29.3 kJ mol⁻¹; Cr(III)–O = 24.5 kJ mol⁻¹] relative to Fe–OH/O [e.g., Fe(III)–O = 23.7 kJ mol⁻¹].⁷⁵ Moreover, very low rates of H₂O exchange of Cr(III) has been suggested to explain the higher resistance of Cr-substituted goethite.⁷⁶ In addition to metal substitution, larger particle sizes and smaller surface areas could explain the slower dissolution rates of natural samples.

We also tested the potential of the selected 24 h hot 6 M HCl treatment in extracting Fe from sheet silicates that could co-occur with the Fe (oxyhydr)oxides in the Ni laterites.^{77,78} Significant amounts of Fe (65–70%) were extracted from nontronite and serpentine. We compared these with the boiling 12 M HCl treatment previously used to extract sheet silicate Fe⁴¹ and documented that the optimized 6 M HCl released 2.5 times higher Fe, most likely due to the higher S/L ratio and longer extraction duration. Overall, this demonstrates the ability of the 6 M HCl treatment to dissolve highly crystalline Fe-bearing phases, such as metal-substituted (oxyhydr)oxides and sheet silicates, without affecting the residual fraction hosted in chromite.

Enhanced Efficiency of the Optimized SEP. The resulting extraction scheme in Figure 3 was validated with mineral mixtures and nickel laterite CRMs (Table 1). Chromium recoveries from mixtures were mostly >89%, matching the results of the single extractions. Only the phosphate step targeting the smallest fraction of Cr showed a lower recovery. Moreover, sequential extraction of the CRMs showed well-targeted crystalline Fe phases (Figure S6).²⁴

Easily Mobilizable to Non-Fe Phase-Bound Fractions. Applying the optimized scheme and three existing SEPs^{23,28,32} (Figure 4a-c) to the PAL laterites yielded very small amounts (<0.1% of total Cr) of exchangeable Cr. Our additional phosphate step leached up to 1% of the total Cr but with notable Fe extraction (up to 2.6%; Figure S7) in the transition and saprolite samples. During this step, we observed colloidal formation likely induced by the interaction of negatively charged phosphate ions and negatively charged surfaces of smectites identified through XRD (Figure S10). Some colloids might have passed through the filters, resulting in the detection of Fe. Another possible explanation is the dissolution of Fe minerals previously reported during phosphate extractions of arsenic, although the mechanism of dissolution remains unclear.^{44,79} In the case of the smectite-free limonite samples from Palawan, and from Zambales and Surigao (Figure S9), where up to 7% of total Cr was recovered in the phosphate step, no such Fe extraction was observed. Thus, care should be taken when interpreting phosphate extracted metals from samples containing clays.

In terms of organic Cr, the optimized SEP extracted comparable concentrations with SEP 2, which also used an oxidant (H_2O_2) after the crystalline Fe phase step but generally in lower amounts than SEP 3, where NaOCl was applied before Fe (oxyhydr)oxide dissolution. SEP 1 extracted the most ($\leq 6.6\%$) due to the ligand-promoted dissolution of Crbearing, poorly crystalline Fe (oxyhydr)oxides by

Environmental Science & Technology



Figure 3. Optimized SEP for Cr in Fe-rich laterites. A solid-to-liquid ratio of 1:100 is applied for all except for organic Cr, where 1:20 was employed. Residual Cr is the difference between the total concentration and the sum of the preceding extractable fractions.

 $Na_4P_2O_7^{80-82}$ This is supported by the anomalously high organic bound Fe ($\leq 14\%$) (Figure S7) that correlates with the increasing trend of poorly crystalline Fe (oxyhydr)oxides down the laterite profile.¹⁸ All other SEPs only extracted <1% of total Fe in this step.

The optimized SEP and SEP 2 leached <1% of Cr using hydroxylamine hydrochloride, which targets the Mn phasebound fraction. Conversely, SEP 3 extracted up to four times more Cr despite using the same reducing reagent. Unlike the two SEPs, the hydroxylamine hydrochloride treatment of SEP 3 is preceded by NaOCl. As previously discussed, NaOCl can oxidize mineral-bound Cr(III) to Cr(VI), promoting adsorption onto Mn (oxyhydr)oxides, which are known adsorbents of Cr(VI) species.⁸³ This indicates that for Cr, applying an oxidant like NaOCl before the mineral dissolution step will not only overestimate organic Cr but also succeeding fraction(s) as a result of the carryover of oxidized Cr.

Fe Phase-Bound to Residual Fractions. Our SEP demonstrated the highest recoveries for Cr and Fe bound to Fe minerals. The 1 M HCl step dissolved increasing Cr (0.3–2%) and Fe (1–6%) fractions from PAL-1 to PAL-3, agreeing well with the increasing amount of poorly crystalline Fe (oxyhydr)oxides.¹⁸ SEP 1 showed the poorest recoveries for both poorly crystalline (<1% Cr, <1.5% Fe) and (<3.5% Cr, <12% Fe) crystalline fractions. Inefficient extraction by

hydroxylamine hydrochloride¹⁸ and the nonselective $Na_4P_2O_7$ step may account for these low recoveries. As discussed earlier, the latter can also retard the dissolution of Fe (oxyhydr)oxides.

Complete dissolution of Fe (oxyhydr)oxides by our optimized 6 M HCl step yielded up to an 11-fold increase in crystalline Fe phase-bound Cr and Fe relative to existing SEPs. Residues after this step (Figure 4d-f) reveal the absence of Fe (oxyhydr)oxides, showing only chromite and quartz signals in XRD and SEM (Figure S11). In the limonite residue where chromite is predominant, we have estimated the maximum extractable Cr (Text S7) and showed that our SEP yields the highest extraction efficiency, closest to the maximum value (Figure 4a) and shows ~85% recovery. In contrast, residues from the two most efficient existing SEPs (2 and 3) show the incomplete dissolution of goethite and hematite, especially in the limonite sample. While our 6 M HCl step recovered 78% of total Fe from PAL-1, both the widely used citratebicarbonate-dithionite and 6 M HCl steps of SEPs 2 and 3, respectively, only recovered half.

Chromium K-edge XAS of the PAL-1 residues unveils how the different SEPs affect the dissolution of Cr. The Cr K-edge HERFD-XANES spectrum of the initial sample shows features common to chromite, Cr-hematite, and Cr-goethite, while our SEP's residue displays features (dotted lines in Figure 4g) analogous only to chromite. Chromium EXAFS fitting of PAL-1 (Figure 4h,i, Table S6) revealed structural incorporation in chromite and Fe (oxyhydr)oxides and resulted in 7.5 neighboring oxygen (O) atoms at 2.00 Å, 4.8 Me₁ at 3.05 Å, 6.0 Me₂ at 3.27 Å, and 7.7 Me₃ at 3.48 Å, where "Me" cations correspond to Cr and Fe. These cations have close atomic numbers and contribute similarly to the EXAFS signal,⁸⁴ and therefore cannot be distinguished from each other. The Cr-O distance is consistent with the octahedral coordination of Cr(III) in chromite and Fe (oxyhydr)oxides but has a slightly higher coordination number (CN) that could be explained by the relatively high value of the correlated Debye-Waller factor (σ^2) and uncertainty (20–25%) of EXAFS CNs.⁸⁵ The 3.05 Å Cr-Me₁ distance is similar but slightly longer than the second shell Cr-Cr/Fe distances of chromite and Cr-Fe (oxyhydr)oxides (2.97-2.98 Å). Rather, it is analogous to the average of the second shell distances and the Cr-Fe_{F2} distance of Crgoethite (3.11 Å). Such strongly overlapping FT peaks often occur in natural heterogeneous samples and complicate the EXAFS fitting.⁸⁶ Similarly, the Cr-Me₂ distance is likely the average of 3.11 Å and the distances of corner-shared Cr-Fe atoms of Cr-goethite and Cr-hematite (3.41-3.44 Å), resulting in a single peak at 3.27 Å. These corner-shared distances of Fe (oxyhydr)oxides and that of chromite (Cr-Fe_c = 3.52 Å) could account for the fourth shell of PAL-1 at 3.48 Å. Only the outermost shell of Cr-hematite (\sim 3.7 Å) could not be fitted in PAL-1. This may be due to the heterogeneity of Cr location as observed in previous EXAFS fitting,⁸⁷ where the outer shells of hematite were removed from the fit of a sample containing both hematite and goethite.

The PAL-1 residues after SEP 2 and 3 were similarly fitted with four shells, while the residue after our SEP was best fitted with only three shells. Cr–Me₁ distances considerably differ and exhibit a decreasing trend from the initial value: SEP 2 (3.02 Å) > SEP 3 (2.99 Å) > optimized SEP (2.96 Å). This indicates decreasing (down to nonexistent) contributions from the Cr–Fe_{E2} shell of Cr-goethite and signifies the dissolution of Cr-bearing Fe (oxyhydr)oxides from the previous extraction



Figure 4. (a-c) Comparison of Cr partitioning (% of total concentration) and (d-f) residual fraction mineralogy of the Palawan Ni laterites (PAL-1—limonite, PAL-2—transition zone, PAL-3—saprolite) based on the optimized method (this study) and existing SEPs (SEP 1–3). The dashed line in (a) represents the maximum extractable Cr further discussed in Text S7. (g) Cr K-edge HERFD-XANES, (h) k^3 -weighted EXAFS spectra, and the corresponding (i) Fourier transforms (FT) of PAL-1 and its SEP residues and reference minerals. Vertical dashed lines in (g) denote features discussed in the text. Dotted lines superimposed on the solid lines in (h) and (i) denote shell-by-shell fits of the EXAFS data, respectively. Fit parameters are given in Table S6. G—goethite, H—hematite, S—spinel, Sm—smectite, Se—serpentine, and Q—quartz.

steps. The Cr–Me₂ shells of SEP 2 (3.23 Å) and SEP 3 (3.20 Å) residues also show a similar trend, while our SEP's residue lacks this atomic correlation. Instead, our SEP's residue has a second shell distance of 3.49 Å and an overall fit matching the local bonding environment of Cr in chromite^{88,89} only (Table S6). Such findings emphasize (1) the importance of laterite Fe (oxyhydr)oxides as hosts for Cr, and (2) previous SEPs only partially dissolve these minerals and underestimate their contributions.

Even in samples with smaller amounts of Fe minerals (PAL-2 and PAL-3), SEPs 2 and 3 exhibited limited dissolution of these phases (Figure 4e,f). On the other hand, our SEP completely dissolved these Fe phases, leaving chromite, quartz, and traces of amorphous silicate, showing a broad XRD peak at

1.57 Å. SEM-EDS confirms the presence of these Cr-free silicate phases (Figure S11b,c). With the optimization of Fe phase-bound extraction steps, our SEP significantly increased the recovery of extractable Cr, providing the best estimate of Cr in the Ni laterites. Additionally, it significantly increased the recovery of other equally important metals such as Mn and Ni (Figure S8).

Further tests on Zambales and Surigao limonites (Figure S9) revealed that our SEP yielded consistent high total recoveries for extractable and non-residual Cr (38-48%) and Fe (82-89%). In these samples, we highlight the importance of including the phosphate step and optimizing the crystalline Fe phase-bound step, especially because these steps extract the dominant pools for Cr. Adsorbed Cr comprised a maximum of

7% of the total Cr, while the crystalline Fe phase-bound fraction comprised the largest non-residual pool for Cr (29-44%) and Fe (82-89%).

Environmental Implications. Existing SEPs applied to tropical soils fail to adequately partition Cr from Fe-rich laterites because they do not consider the different species of Cr and its ability to stabilize the crystal structure of Fe (oxyhydr)oxides. By robust calibration with appropriate mineral standards, we optimized and validated a SEP (Figure 3) that is reproducible, efficient, and selective for Cr in such Fe-rich materials. Identifying the binding sites of Cr is crucial for assessing its bioavailability, potential mobility, and transport mechanisms from laterites all the way to drinking and groundwaters. Our newly optimized SEP offers an important tool for monitoring and predicting pathways for Cr release that could ensue due to changes in environmental conditions (e.g., pH, redox, etc.). Our results also point to possible best practices for managing Cr-rich laterites, where the mobilization of Cr through weathering and mining may be linked to downstream containment and remediation efforts.

Exchangeable Cr targeted by $Ca(NO_3)_2$ represents easily mobilizable Cr in the presence of elevated salt inputs such as during saltwater interaction⁹⁰ or irrigation.⁹¹ Adsorbed Cr could be liberated by phosphorus sources such as agricultural drainage.⁹⁰ In a study by Becquer et al.,¹⁶ increased Cr concentrations in soil solutions was correlated to the desorption of Cr(VI) by phosphorus fertilizer inputs. Thus, accounting for adsorbed Cr is vital, especially in areas affected by agriculture and rehabilitation in the case of mining areas.

Chromium incorporated in Mn- and Fe-phases are more conservative pools and were only leached by reductive dissolution (NH₂OH·HCl) and protonation (HCl). Thus, potential Cr release may occur under reducing (e.g., by bacterial activity)⁹¹ and acidic conditions (e.g., by organic acids). For example, common organic acids, such as oxalate and citrate, have been found to solubilize Cr-bearing goethite.⁹² Reducing and acidic conditions are also used in the hydrometallurgical processing of laterites (e.g., reductive bioleaching⁹³ and high-pressure acid leaching⁹⁴). In such cases, Mn and predominant Fe (oxyhydr)oxides are dissolved to solubilize associated Ni and Co, and knowledge of the extractable Cr from these phases is crucial in monitoring downstream processes and developing strategies for the immobilization of the leached Cr. Moreover, we demonstrated that interaction with strong oxidants such as NaOCl could potentially release Cr from Cr-bearing Fe (oxyhydr)oxides. NaOCl is extensively used in water treatment and has been found to oxidize Cr(III) to Cr(VI) during chlorination of drinking water.⁶⁷ Our results warrant further research to assess the occurrence of Cr during drinking water treatment in lateritic areas. Furthermore, we were able to distinguish the extractable fractions from the residual chromite-bound Cr, which represents the weathering-resistant and most stable pool for Cr.

Using our optimized SEP, we demonstrated through the example of Philippine Ni laterites that these Fe-rich materials are significant sources of easily mobilizable and toxic Cr(VI) comprising up to 7% of total Cr. These fractions correspond to $30-1192 \text{ mg kg}^{-1}$ Cr(VI) and are comparable to Cr(VI) detected in laterites from New Caledonia (\leq 358 mg kg⁻¹)¹⁹ and Brazil (\leq 1014 mg kg⁻¹).³⁰ We also highlight the predominant association and structural incorporation of Cr as Cr(III) in Fe (oxyhydr)oxides, suggesting a potential release

of Cr during the hydrometallurgical processing of laterites. The quantification of these important Cr reservoirs is crucial in ensuring the meticulous and sustainable management of laterite mining and processing regions.

ASSOCIATED CONTENT

Supporting Information

The Supporting Information is available free of charge at https://pubs.acs.org/doi/10.1021/acs.est.3c10774.

Methods and mineral characterization; elemental composition of mineral standards; extraction protocols; dissolution efficiencies and Cr recoveries of mineral standards and laterites; EXAFS fitting results; and supporting XRD, FTIR, SEM data, and micrographs (PDF)

AUTHOR INFORMATION

Corresponding Author

Ruth Esther G. Delina – GFZ German Research Centre for Geosciences, Telegrafenberg, 14473 Potsdam, Germany; Department of Earth Sciences, Freie Universität Berlin, 12249 Berlin, Germany; orcid.org/0009-0000-1895-1937; Email: rdelina@gfz-potsdam.de

Authors

- Jeffrey Paulo H. Perez GFZ German Research Centre for Geosciences, Telegrafenberg, 14473 Potsdam, Germany; orcid.org/0000-0002-0256-0576
- Jessica A. Stammeier GFZ German Research Centre for Geosciences, Telegrafenberg, 14473 Potsdam, Germany
- Elena F. Bazarkina The Rossendorf Beamline at ESRF, The European Synchrotron, 38043 Grenoble Cedex 9, France; Institute of Resource Ecology, Helmholtz-Zentrum Dresden-Rossendorf, 01328 Dresden, Germany
- Liane G. Benning GFZ German Research Centre for Geosciences, Telegrafenberg, 14473 Potsdam, Germany; Department of Earth Sciences, Freie Universität Berlin, 12249 Berlin, Germany; orcid.org/0000-0001-9972-5578

Complete contact information is available at: https://pubs.acs.org/10.1021/acs.est.3c10774

Notes

The authors declare no competing financial interest.

ACKNOWLEDGMENTS

R.E.G. Delina acknowledges the DAAD German Academic Exchange Service for her Doctoral Research Scholarship (grant no. 91769931). J.P.H. Perez is in part funded by his independent research fellowship (GFZ Discovery Fund; grant no. P-032-45-002). This project has received funding from the Helmholtz Recruiting Initiative, awarded to L.G. Benning (award no. I-044-16-01). We acknowledge the European Synchrotron Radiation Facility for the provision of beamtime on beamline BM20 (proposal EV-495) and thank K.O. Kvashnina and A.C. Scheinost for assistance. All ICP-OES analyses were performed at the Elements and Minerals of the Earth Laboratory with the assistance of S. Tonn and the Helmholtz Laboratory for the Geochemistry of the Earth Surface at GFZ Potsdam. We thank R. Blukis for the help during the Rietveld refinements and R. Trumbull for providing us with the chromite reference.

Environmental Science & Technology

REFERENCES

(1) Stoops, G.; Marcelino, V., Chapter 24 – Lateritic and Bauxitic Materials. In *Interpretation of Micromorphological Features of Soils and Regoliths*, 2nd ed.; Stoops, G.; Marcelino, V.; Mees, F., Eds.; Elsevier, 2018; pp 691–720.

(2) Tardy, Y. Pétrologie des latérites et des sols tropicaux; Masson: Paris, 1993, p 459.

(3) Elias, M. Nickel laterite deposits - geological overview, resources and exploitation. *CODES Spec. Publ.* **2002**, *4*, 205–220.

(4) Butt, C. R. M.; Cluzel, D. Nickel Laterite Ore Deposits: Weathered Serpentinites. *Elements* 2013, 9 (2), 123–128.

(5) Aiglsperger, T.; Proenza, J. A.; Lewis, J. F.; Labrador, M.; Svojtka, M.; Rojas-Purón, A.; Longo, F.; Ďurišová, J. Critical metals (REE, Sc, PGE) in Ni laterites from Cuba and the Dominican Republic. *Ore Geol. Rev.* **2016**, *73*, 127–147.

(6) USGS. *Mineral Commodity Summaries* 2023; United States Geological Survey, 2023, p 210.

(7) Chrysochoou, M.; Theologou, E.; Bompoti, N.; Dermatas, D.; Panagiotakis, I. Occurrence, Origin and Transformation Processes of Geogenic Chromium in Soils and Sediments. *Curr. Pollut. Rep.* **2016**, 2 (4), 224–235.

(8) Ulrich, M.; Cathelineau, M.; Muñoz, M.; Boiron, M.-C.; Teitler, Y.; Karpoff, A. M. The relative distribution of critical (Sc, REE) and transition metals (Ni, Co, Cr, Mn, V) in some Ni-laterite deposits of New Caledonia. *J. Geochem. Explor.* **2019**, *197*, 93–113.

(9) Wedepohl, K. H. The composition of the continental crust. Geochim. Cosmochim. Acta 1995, 59 (7), 1217-1232.

(10) Guertin, J.; Jacobs, J. A.; Avakian, C. P. Chromium(VI) Handbook; CRC Press, 2016, p 761.

(11) Katz, S.; Salem, H. The Biological and Environmental Chemistry of Chromium; VCH: USA, 1994, p 220.

(12) WHO. Chromium in Drinking-Water: Background Document for Development of WHO Guidelines for Drinking-Water Quality; World Health Organization: Switzerland, 2003.

(13) Oze, C.; Fendorf, S.; Bird, D. K.; Coleman, R. G. Chromium geochemistry in serpentinized ultramafic rocks and serpentine soils from the Franciscan complex of California. *Am. J. Sci.* **2004**, *304* (1), 67–101.

(14) Kotaś, J.; Stasicka, Z. Chromium occurrence in the environment and methods of its speciation. *Environ. Pollut.* **2000**, *107* (3), 263–283.

(15) Richard, F. C.; Bourg, A. C. M. Aqueous geochemistry of chromium: A review. *Water Res.* **1991**, 25 (7), 807–816.

(16) Becquer, T.; Quantin, C.; Sicot, M.; Boudot, J. P. Chromium availability in ultramafic soils from New Caledonia. *Sci. Total Environ.* **2003**, *301* (1–3), 251–261.

(17) Economou-Eliopoulos, M.; Frei, R.; Megremi, I. Potential leaching of Cr(VI) from laterite mines and residues of metallurgical products (red mud and slag): An integrated approach. *J. Geochem. Explor.* **2016**, *162*, 40–49.

(18) Delina, R. E.; Arcilla, C.; Otake, T.; Garcia, J. J.; Tan, M.; Ito, A. Chromium occurrence in a nickel laterite profile and its implications to surrounding surface waters. *Chem. Geol.* **2020**, *558*, 119863.

(19) Gunkel-Grillon, P.; Laporte-Magoni, C.; Lemestre, M.; Bazire, N. Toxic chromium release from nickel mining sediments in surface waters, New Caledonia. *Environ. Chem. Lett.* **2014**, *12* (4), 511–516. (20) USEPA. 2006 Edition of the Drinking Water Standards and

Health Advisories; United States Environmental Protection Agency: Washington, DC, USA, 2006.

(21) WHO. Guidelines for Drinking-Water Quality: Fourth Edition Incorporating the First Addendum; World Health Organization: Geneva, Switzerland, 2017.

(22) Gleyzes, C.; Tellier, S.; Astruc, M. Fractionation studies of trace elements in contaminated soils and sediments: a review of sequential extraction procedures. *TrAC, Trends Anal. Chem.* **2002**, *21* (6–7), 451–467.

(23) Silveira, M. L.; Alleoni, L. R.; O'Connor, G. A.; Chang, A. C. Heavy metal sequential extraction methods-a modification for tropical soils. *Chemosphere* **2006**, *64* (11), 1929–1938.

(24) Rodgers, K. J.; Hursthouse, A.; Cuthbert, S. The Potential of Sequential Extraction in the Characterisation and Management of Wastes from Steel Processing: A Prospective Review. *Int. J. Environ. Res. Public Health* **2015**, *12* (9), 11724–11755.

(25) Hass, A.; Fine, P. Sequential Selective Extraction Procedures for the Study of Heavy Metals in Soils, Sediments, and Waste Materials – a Critical Review. *Crit. Rev. Environ. Sci. Technol.* **2010**, 40 (5), 365–399.

(26) Bacon, J. R.; Davidson, C. M. Is there a future for sequential chemical extraction? *Analyst* **2008**, *133* (1), 25–46.

(27) Becquer, T.; Quantin, C.; Rotte-Capet, S.; Ghanbaja, J.; Mustin, C.; Herbillon, A. J. Sources of trace metals in Ferralsols in New Caledonia. *Eur. J. Soil Sci.* **2006**, *57* (2), 200–213.

(28) Quantin, C.; Becquer, T.; Rouiller, J.; Berthelin, J. Redistribution of Metals in a New Caledonia Ferralsol After Microbial Weathering. *Soil Sci. Soc. Am. J.* **2002**, *66* (6), 1797–1804.

(29) Schwertmann, U. Solubility and dissolution of iron oxides. *Plant Soil* **1991**, *130* (1–2), 1–25.

(30) Garnier, J.; Quantin, C.; Martins, E. S.; Becquer, T. Solid speciation and availability of chromium in ultramafic soils from Niquelândia (Brazil): chemical and spectroscopic approaches. *J. Geochem. Explor.* **2006**, 88 (1–3), 206–209.

(31) Tashakor, M.; Zuhairi Wan Yaacob, W.; Mohamad, H.; Abdul Ghani, A.; Saadati, N. Assessment of selected sequential extraction and the toxicity characteristic leaching test as indices of metal mobility in serpentinite soils. *Chem. Speciation Bioavailability* **2014**, *26* (3), 139–147.

(32) Benitez, L. N.; Dubois, J.-P. Evaluation of the Selectivity of Sequential Extraction Procedures Applied to the Speciation of Cadmium in Soils. *Int. J. Environ. Anal. Chem.* **1999**, 74 (1–4), 289–303.

(33) Doelsch, E.; Moussard, G.; Macary, H. S. Fractionation of tropical soilborne heavy metals–Comparison of two sequential extraction procedures. *Geoderma* **2008**, *143* (1-2), 168–179.

(34) Tessier, A.; Campbell, P. G. C.; Bisson, M. Sequential extraction procedure for the speciation of particulate trace metals. *Anal. Chem.* **1979**, *51* (7), 844–851.

(35) Leleyter, L.; Probst, J.-L. A New Sequential Extraction Procedure for the Speciation of Particulate Trace Elements in River Sediments. *Int. J. Environ. Anal. Chem.* **1999**, 73 (2), 109–128.

(36) Shuman, L. M. Fractionation method for soil microelements. *Soil Sci.* **1985**, *140* (1), *11–22*.

(37) Arcilla, C.; Reyes, R.; Aguda, N.; Panlaqui, A.; Ruelo, R.; Vargas, J.; Palattao, B. *Exploratory Characterization of Scandium and Rare Earth Elements on Zambales, Palawan and Surigao Del Norte Nickeliferous Laterite Deposits*; Department of Science and Technology Philippine Council for Industry, Energy, and Emerging Technology Research and Development (DOST-PCIEERD), 2019, p 131.

(38) Fandeur, D.; Juillot, F.; Morin, G.; Olivi, L.; Cognigni, A.; Ambrosi, J.-P.; Guyot, F.; Fritsch, E. Synchrotron-based speciation of chromium in an Oxisol from New Caledonia: Importance of secondary Fe-oxyhydroxides. *Am. Mineral.* 2009, 94 (5–6), 710–719.
(39) Schwertmann, U.; Cornell, R. M. *Iron Oxides in the Laboratory: Preparation and Characterization*; Wiley: Germany, 2000, p 204.

(40) Filgueiras, A. V.; Lavilla, I.; Bendicho, C. Chemical sequential extraction for metal partitioning in environmental solid samples. *Journal of Environmental Monitoring* **2002**, *4* (6), 823–857.

(41) Poulton, S. W.; Canfield, D. E. Development of a sequential extraction procedure for iron: implications for iron partitioning in continentally derived particulates. *Chem. Geol.* **2005**, *214* (3–4), 209–221.

(42) Raiswell, R.; Canfield, D. E.; Berner, R. A. A comparison of iron extraction methods for the determination of degree of pyritisation and the recognition of iron-limited pyrite formation. *Chem. Geol.* **1994**, *111* (1–4), 101–110.

(43) Voelz, J. L.; Johnson, N. W.; Chun, C. L.; Arnold, W. A.; Penn, R. L. Quantitative Dissolution of Environmentally Accessible Iron Residing in Iron-Rich Minerals: A Review. *ACS Earth Space Chem.* **2019**, *3* (8), 1371–1392.

(44) Drahota, P.; Grosslova, Z.; Kindlova, H. Selectivity assessment of an arsenic sequential extraction procedure for evaluating mobility in mine wastes. *Anal. Chim. Acta* **2014**, *839*, 34–43.

(45) Tokunaga, T. K.; Lipton, D. S.; Benson, S. M.; Yee, A. W.; Oldfather, J. M.; Duckart, E. C.; Johannis, P. W.; Halvorsen, K. E. Soil selenium fractionation, depth profiles and time trends in a vegetated site at Kesterson Reservoir. *Water, Air, Soil Pollut.* **1991**, 57–58 (1), 31–41.

(46) La Force, M. J.; Fendorf, S. Solid-Phase Iron Characterization During Common Selective Sequential Extractions. *Soil Sci. Soc. Am. J.* **2000**, *64* (5), 1608–1615.

(47) Shuman, L. M. Sodium Hypochlorite Methods for Extracting Microelements Associated with Soil Organic Matter. *Soil Sci. Soc. Am. J.* **1983**, 47 (4), 656–660.

(48) Denys, A.; Janots, E.; Auzende, A.-L.; Lanson, M.; Findling, N.; Trcera, N. Evaluation of selectivity of sequential extraction procedure applied to REE speciation in laterite. *Chem. Geol.* **2021**, *559*, 119954.

(49) Claff, S. R.; Sullivan, L. A.; Burton, E. D.; Bush, R. T. A sequential extraction procedure for acid sulfate soils: Partitioning of iron. *Geoderma* **2010**, *155* (3–4), 224–230.

(50) Porsch, K.; Kappler, A. FeII oxidation by molecular O2 during HCl extraction. *Environ. Chem.* **2011**, *8* (2), 190–197.

(51) Bokhari, S. N. H.; Meisel, T. C. Method Development and Optimisation of Sodium Peroxide Sintering for Geological Samples. *Geostand. Geoanal. Res.* **2017**, *41* (2), 181–195.

(52) Bousserrhine, N.; Gasser, U. G.; Jeanroy, E.; Berthelin, J. Bacterial and Chemical Reductive Dissolution of Mn-Co-Cr-and Al-Substituted Goethites. *Geomicrobiol. J.* **1999**, *16* (3), 245–258.

(53) Hua, J.; Chen, M.; Liu, C.; Li, F.; Long, J.; Gao, T.; Wu, F.; Lei, J.; Gu, M. Cr Release from Cr-Substituted Goethite during Aqueous Fe(II)-Induced Recrystallization. *Minerals* **2018**, *8* (9), 367.

(54) Li, M.; Liu, H.; Chen, T.; Wei, L.; Wang, C.; Hu, W.; Wang, H. The transformation of α -(Al, Fe)OOH in natural fire: Effect of Al substitution amount on fixation of phosphate. *Chem. Geol.* **2019**, *524*, 368–382.

(55) Trolard, F.; Bourrie, G.; Jeanroy, E.; Herbillon, A. J.; Martin, H. Trace metals in natural iron oxides from laterites: A study using selective kinetic extraction. *Geochim. Cosmochim. Acta* **1995**, *59* (7), 1285–1297.

(56) Ekstrom, E. B.; Learman, D. R.; Madden, A. S.; Hansel, C. M. Contrasting effects of Al substitution on microbial reduction of Fe(III) (hydr)oxides. *Geochim. Cosmochim. Acta* **2010**, 74 (24), 7086–7099.

(57) Kaur, N.; Gräfe, M.; Singh, B.; Kennedy, B. Simultaneous Incorporation of Cr, Zn, Cd, and Pb in the Goethite Structure. *Clays Clay Miner.* **2009**, *57* (2), 234–250.

(58) Dublet, G.; Juillot, F.; Brest, J.; Noël, V.; Fritsch, E.; Proux, O.; Olivi, L.; Ploquin, F.; Morin, G. Vertical changes of the Co and Mn speciation along a lateritic regolith developed on peridotites (New Caledonia). *Geochim. Cosmochim. Acta* **2017**, *217*, 1–15.

(59) Ugwu, I. M.; Sherman, D. M. The solubility of goethite with structurally incorporated nickel and cobalt: Implication for laterites. *Chem. Geol.* **2019**, *518*, 1–8.

(60) Cornell, R. M.; Schwertmann, U. The Iron Oxides: Structure, Properties, Reactions, Occurrences and Uses; Wiley: Germany, 2003, p 664.

(61) Biber, M. V.; dos Santos Afonso, M.; Stumm, W. The coordination chemistry of weathering: IV. Inhibition of the dissolution of oxide minerals. *Geochim. Cosmochim. Acta* **1994**, 58 (9), 1999–2010.

(62) Kraal, P.; van Genuchten, C. M.; Behrends, T.; Rose, A. L. Sorption of phosphate and silicate alters dissolution kinetics of poorly crystalline iron (oxyhydr)oxide. *Chemosphere* **2019**, 234, 690–701.

(63) Majzlan, J. Thermodynamic Stabilization of Hydrous Ferric Oxide by Adsorption of Phosphate and Arsenate. *Environ. Sci. Technol.* **2011**, 45 (11), 4726–4732.

(64) Ruttenberg, K. C. Development of a sequential extraction method for different forms of phosphorus in marine sediments. *Limnol. Oceanogr.* **1992**, 37 (7), 1460–1482.

(65) Mikutta, R.; Kleber, M.; Kaiser, K.; Jahn, R. Review. Soil Sci. Soc. Am. J. 2005, 69 (1), 120–135.

(66) Lee, G.; Hering, J. G. Oxidative Dissolution of Chromium(III) Hydroxide at pH 9, 3, and 2 with Product Inhibition at pH 2. *Environ. Sci. Technol.* **2005**, *39* (13), 4921–4928.

(67) Lindsay, D. R.; Farley, K. J.; Carbonaro, R. F. Oxidation of CrIII to CrVI during chlorination of drinking water. *J. Environ. Monit.* **2012**, *14* (7), 1789–1797.

(68) Gruebel, K. A.; Davis, J. A.; Leckie, J. O. The Feasibility of Using Sequential Extraction Techniques for Arsenic and Selenium in Soils and Sediments. *Soil Sci. Soc. Am. J.* **1988**, *52* (2), 390–397.

(69) Wright, M. T.; Parker, D. R.; Amrhein, C. Critical Evaluation of the Ability of Sequential Extraction Procedures To Quantify Discrete Forms of Selenium in Sediments and Soils. *Environ. Sci. Technol.* **2003**, 37 (20), 4709–4716.

(70) Kraal, P.; van Genuchten, C. M.; Behrends, T. Phosphate coprecipitation affects reactivity of iron (oxyhydr)oxides towards dissolved iron and sulfide. *Geochim. Cosmochim. Acta* **2022**, 321, 311–328.

(71) Muehe, E. M.; Adaktylou, I. J.; Obst, M.; Zeitvogel, F.; Behrens, S.; Planer-Friedrich, B.; Kraemer, U.; Kappler, A. Organic Carbon and Reducing Conditions Lead to Cadmium Immobilization by Secondary Fe Mineral Formation in a pH-Neutral Soil. *Environ. Sci. Technol.* **2013**, 47 (23), 13430–13439.

(72) Patzner, M. S.; Mueller, C. W.; Malusova, M.; Baur, M.; Nikeleit, V.; Scholten, T.; Hoeschen, C.; Byrne, J. M.; Borch, T.; Kappler, A.; Bryce, C. Iron mineral dissolution releases iron and associated organic carbon during permafrost thaw. *Nat. Commun.* **2020**, *11* (1), 6329.

(73) Alvarez, M.; Rueda, E. H.; Sileo, E. E. Simultaneous incorporation of Mn and Al in the goethite structure. *Geochim. Cosmochim. Acta* **2007**, *71* (4), 1009–1020.

(74) Lim-Nunez, R.; Gilkes, R. J. Acid dissolution of synthetic metalcontaining goethites and hematites In *Proceedings of the International Clay Conference Denver, 1985*; Clay Minerals Society, 1985, pp 197– 204.

(75) Weast, R. C. CRC Handbook of Chemistry and Physics, 1st student ed ed.; CRC Press: USA, 1988; p 1760.

(76) Burgess, J. Ions in Solution: Basic Principles of Chemical Interactions; Halsted Press: USA, 1988, p 191.

(77) Elias, M.; Donaldson, M. J.; Giorgetta, N. E. Geology, mineralogy, and chemistry of lateritic nickel-cobalt deposits near Kalgoorlie, Western Australia. *Econ. Geol.* **1981**, *76* (6), 1775–1783.

(78) Marsh, E. E.; Anderson, E. D.; Gray, F. Nickel-cobalt Laterites – A Deposit Model, Chapter H of Mineral Deposit Models for Resource Assessment; U.S. Geological Survey Scientific Investigations Report 2010-5070-H: Reston, VA, 2013, p 38.

(79) Eiche, E.; Kramar, U.; Berg, M.; Berner, Z.; Norra, S.; Neumann, T. Geochemical changes in individual sediment grains during sequential arsenic extractions. *Water Res.* **2010**, *44* (19), 5545–5555.

(80) Kaiser, K.; Zech, W. Defects in Estimation of Aluminum in Humus Complexes of Podzolic Soils by Pyrophosphate Extraction. *Soil Sci.* **1996**, *161* (7), 452–458.

(81) Kaiser, M.; Ellerbrock, R. H.; Wulf, M.; Dultz, S.; Hierath, C.; Sommer, M. The influence of mineral characteristics on organic matter content, composition, and stability of topsoils under long-term arable and forest land use. *J. Geophys. Res.: Biogeosci.* **2012**, *117* (G2), 1–16.

(82) Shuman, L. M. Separating Soil Iron- and Manganese-Oxide Fractions for Microelement Analysis. *Soil Sci. Soc. Am. J.* **1982**, *46* (5), 1099–1102.

6400

(83) Islam, M. A.; Angove, M. J.; Morton, D. W.; Pramanik, B. K.; Awual, M. R. A mechanistic approach of chromium (VI) adsorption onto manganese oxides and boehmite. *J. Environ. Chem. Eng.* **2020**, 8 (2), 103515.

(84) Dublet, G.; Juillot, F.; Morin, G.; Fritsch, E.; Fandeur, D.; Ona-Nguema, G.; Brown, G. E. Ni speciation in a New Caledonian lateritic regolith: A quantitative X-ray absorption spectroscopy investigation. *Geochim. Cosmochim. Acta* **2012**, *95*, 119–133.

(85) Penner-Hahn, J. E. Characterization of "spectroscopically quiet" metals in biology. *Coord. Chem. Rev.* **2005**, 249 (1–2), 161–177.

(86) Foster, A. L.; Brown, G. E.; Tingle, T. N.; Parks, G. A. Quantitative arsenic speciation in mine tailings using X-ray absorption spectroscopy. *Am. Mineral.* **1998**, 83 (5–6), 553–568.

(87) Marshall, T. A.; Morris, K.; Law, G. T.; Livens, F. R.; Mosselmans, J. F.; Bots, P.; Shaw, S. Incorporation of Uranium into Hematite during crystallization from ferrihydrite. *Environ. Sci. Technol.* **2014**, 48 (7), 3724–3731.

(88) Galivarapu, J. K.; Kumar, D.; Banerjee, A.; Sathe, V.; Aquilanti, G.; Rath, C. Effect of size reduction on cation distribution and magnetic transitions in $CoCr_2O_4$ multiferroic: EXAFS, magnetic and diffused neutron scattering measurements. *RSC Adv.* **2016**, *6* (68), 63809–63819.

(89) Peterson, M. L.; Brown, G. E.; Parks, G. A.; Stein, C. L. Differential redox and sorption of Cr (III/VI) on natural silicate and oxide minerals: EXAFS and XANES results. *Geochim. Cosmochim. Acta* **1997**, *61* (16), 3399–3412.

(90) Keon, N. E.; Swartz, C. H.; Brabander, D. J.; Harvey, C.; Hemond, H. F. Validation of an Arsenic Sequential Extraction Method for Evaluating Mobility in Sediments. *Environ. Sci. Technol.* **2001**, 35 (13), 2778–2784.

(91) Pickering, W. F. Metal ion speciation – soils and sediments (a review). Ore Geol. Rev. **1986**, 1 (1), 83–146.

(92) Sun, S.; Deng, T.; Ao, M.; Mo, Y.; Li, J.; Liu, T.; Yang, W.; Jin, C.; Qiu, R.; Tang, Y. Release of chromium from Cr(III)- and Ni(II)substituted goethite in presence of organic acids: Role of pH in the formation of colloids and complexes. *Sci. Total Environ.* **2023**, *904*, 166979.

(93) Johnson, D. B.; Smith, S. L.; Santos, A. L. Bioleaching of Transition Metals From Limonitic Laterite Deposits and Reassessment of the Multiple Roles of Sulfur-Oxidizing Acidophiles in the Process. *Front. Microbiol.* **2021**, *12*, 703177.

(94) Whittington, B. I.; Muir, D. Pressure Acid Leaching of Nickel Laterites: A Review. *Miner. Process. Extr. Metall. Rev.* **2000**, *21* (6), 527–599.